

Modified blue TiO₂ nanostructures for efficient photo-oxidative removal of harmful NO_x gases

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Abstract—Blue TiO₂ nanostructures were produced via Lithium/ethylenediamine (Li/EDA) reduction method and applied for photo-oxidative removal of harmful NO_x gases under simulated solar light irradiation. Blue TiO₂ possesses some unique physicochemical properties such as enhanced visible-light absorption, superficial defects or oxygen vacancies, and the evolution of Ti³⁺ species. Moreover, the photoluminescence spectra (PL) revealed the efficient separation of photoinduced electron-hole pairs in the modified blue TiO₂ nanostructures, enhancing their photocatalytic activities. The results indicated that the blue TiO₂ nanostructures exhibited the highest performance towards photo-oxidation of NO_x gases, with an efficiency of 72.6% under simulated solar light irradiation.

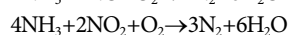
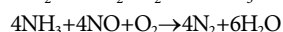
Keywords: Photocatalyst, Blue TiO₂, Oxygen Vacancy, NO_x Gas

INTRODUCTION

Air pollution is fully acknowledged as a major environmental problem owing to its direct influence on the ecosystem and all living organisms. Exhaust gases after burning fossil fuels are the main contributors to air pollution. These gases are mainly composed of carbon monoxide (CO), carbon dioxide (CO₂), sulfur dioxide (SO₂), and the different forms of nitrogen oxides (NO_x) [1]. Among them, NO_x, which refers mainly to a mixture of NO and NO₂, has been paid much attention to scientists due to adverse effects on the environment and living organisms. The emission of NO_x in the atmosphere is responsible for serious environmental concerns, including acid rain, photochemical smog, and stratospheric ozone depletion [1-3]. Moreover, NO_x gases affect the respiratory system, causing inflammation in the respiratory tracks and lungs. Also, long-term exposure to NO_x can reduce lung function and increase the risk of respiratory disease [4,5]. In addition, NO_x gas has revealed that it can have a harmful effect on the physiological metabolism and content of secondary metabolites in plants, causing oxidative damage to nucleic acids, proteins, lipids, aquatic, and terrestrial ecosystems [6]. Due to various hazardous impacts related to NO_x pollution, it is necessary to disinfect or control the released NO_x gases. Currently, one of the effective methods applied to NO_x abatement is preventing the formation of NO_x in the air by using a low-level nitrogen fuel such as natural gas instead of coal or heavy oil [1]. However, natural gas cannot replace high nitrogen fuels in all situations.

Decomposition of the NO_x gas by using a selective non-catalytic

reduction process could be an option. In this process, ammonia (NH₃) or urea (NH₂CONH₂) is injected into the firebox of the boiler at high temperature to react with the nitrogen oxides formed in the combustion process [7,8]. Because of the chemical redox reaction, nitrogen gas is produced instead of NO_x based on the following possible reactions:



Although the selective non-catalytic reduction method has many advantages, such as relatively simple installation, low investment costs, low energy use, and little space occupied, the process has some drawbacks too. The requirement of high temperature, the optimum reaction temperature in a narrow range, and secondary emission of gases like CO₂ and NH₃ are challenging. These negatives are limiting commercial applications of the selective non-catalytic reduction process.

In the last decades, one of the promising techniques to control environmental pollution is the use of photocatalysts, which can efficiently convert these pollutants into harmless products by utilizing green solar energy [9]. TiO₂ is a typical standard photocatalyst because of its unique properties, such as high photo-responsivity, non-toxic, and stable photocatalyst. Unfortunately, TiO₂ can only be excited under ultraviolet light (UV) to be active for photocatalysis due to its relatively larger band gap energy (~3.2 eV for anatase phase and ~3.0 eV of rutile phase) [10]. Thus, these drawbacks of pristine TiO₂ prevent the effective utilization of sunlight. Sunlight comprises about 5% of the UV spectrum, whereas the remaining major portion of the solar spectrum is comprised of the visible spectrum. Therefore, it is necessary to reduce the band gap energy of TiO₂ so

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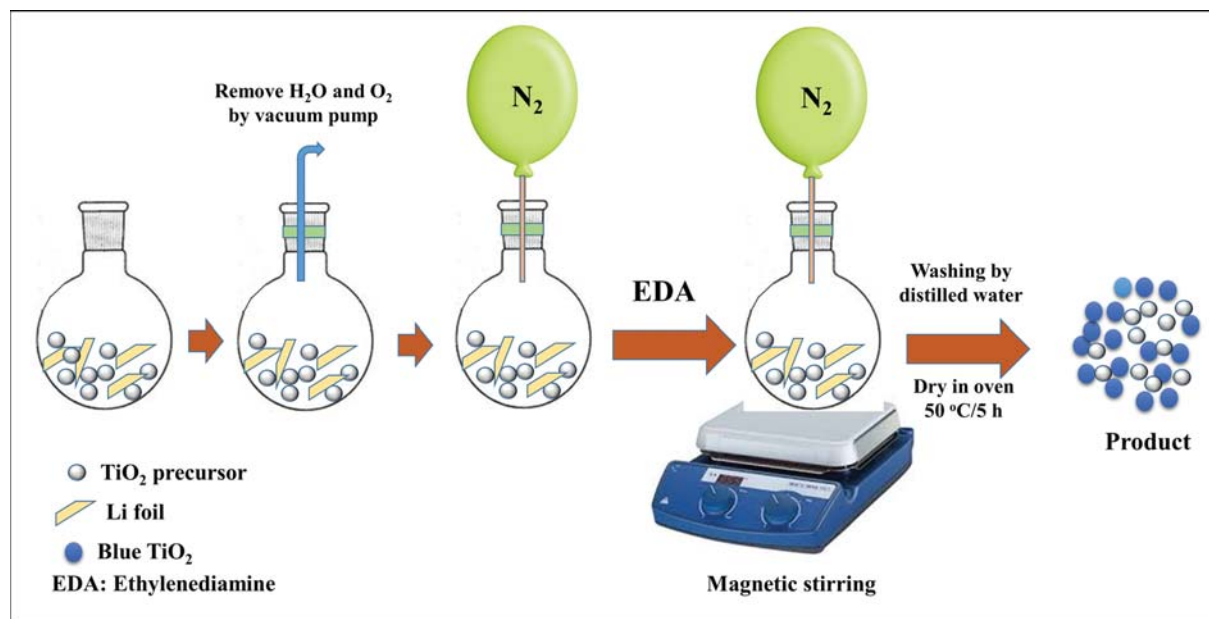


Fig. 1. Synthesis process of b-TiO₂ and n/b-TiO₂.

that the photo-responsive ability can be tuned towards visible light. Blue TiO₂ or surface-disordered TiO₂, which is composed of a thin disordered shell and a lattice core, has received much attention owing to its visible light-responsive characteristics and enhanced photocatalytic activities [11]. Zhang et al. [11] indicated a strong redshift in the absorption to the visible light region of the blue TiO₂. Similarly, Adriana et al. [9] reported broad absorption at wavelengths up to 800 nm, whereas the precursor TiO₂ presents absorption only in the UV light. Moreover, the synthesis process of the blue TiO₂ produces superficial defects, including Ti³⁺ and oxygen vacancies, acting as barriers for the recombination of electron-hole pairs. In addition to the photocatalysis, the blue TiO₂ has also shown improved performance in the dye-sensitized solar cells [12,13], Lithium-ion rechargeable batteries [14], disinfection of freshwater algae [15], and supercapacitors [16,17]. However, a detailed study of surface disordered blue TiO₂ and their different nanostructures for the NO_x removal is still lacking. Hence, in this work, we report the investigation of the reduced blue TiO₂ materials for NO_x removal efficiencies.

EXPERIMENTAL

1. Materials

The precursor TiO₂-P25 (P25) (Degussa), Lithium foil (98%, Samchun Chemical), Ethylenediamine (98%, Sigma Aldrich), H₂SO₄ (Samchun Chemical), HCl (35%, Duksan), and NaOH (Sigma-Aldrich) were used to synthesize blue TiO₂ particles (b-TiO₂), nanostructured TiO₂ (n-TiO₂), and nanostructured blue TiO₂ (n/b-TiO₂).

2. Synthesis of the Reduced Forms of TiO₂

2-1. Synthesis of n-TiO₂

The n-TiO₂ was prepared via a microwave-assisted hydrothermal method as described in our previous work with a minor modification [12]. Briefly, a mixture of 0.5 g of P25 and 25 mL of 10 M

NaOH was stirred gently for 30 min to form a suspension at room temperature. Then, the resulting suspension was transferred to a Teflon vessel reactor for the microwave hydrothermal process. The n-TiO₂ was formed in the microwave reaction system at 100 °C, 295 W, and 300 rpm for 4 h. After that, the n-TiO₂ was washed several times with double distilled water and 0.5 M HCl, finally dried in a freeze dryer (EYELA, FDU-2100) at -80 °C for 20 h. The nanostructured TiO₂ consisted of a mixture of nanoparticles, nanorods, nanotubes, and nanosheets as well.

2-2. Synthesis of b-TiO₂ and n/b-TiO₂

The reduced b-TiO₂ or n/b-TiO₂ were synthesized by reducing P25 or n-TiO₂ using Li/EDA as a reducing agent [9,11-13], according to the process demonstrated in Fig. 1. First, a mixture of 0.35 g of Li foil, 50 mL of EDA, and 0.5 g of precursor TiO₂ (P25 or n-TiO₂) was placed in a glass reactor filled full with nitrogen gas. Then, the mixture was stirred in the dark condition for six days. After that, the obtained product was neutralized by HCl and washed with distilled water and dried in an oven at 40 °C for 6 h. Finally, the powder was ground in a mortar to form b-TiO₂ or n/b-TiO₂.

3. Characterization

The crystal structures of all samples were identified by X-ray diffraction analysis, measured at 2θ angles from 5° to 70° using a monochromatized Cu-Kα (λ=1.54 Å) radiation. The morphologies, microstructures, and particle size of products were observed by high-resolution transmission electron microscopic technique (HRTEM, FEI, TitanTM 80-300). Photoluminescence (PL) spectra were measured using a photoluminescence spectrometer (FLS980, Edinburgh Instruments). X-ray photoelectron spectroscopy (XPS) was measured with an X-ray monochromator, XR5 Gun-500 μm, from Thermo-scientific Instruments. The light absorption ability and band gap values were analyzed by UV-Vis diffuse reflectance spectrophotometer (DRS, UV-Vis V-500, Jasco). The specific surface area of the samples was recorded by N₂ adsorption-desorption mea-

surement via Brunauer Emmett Teller (BET) method using a Micromeritics ASAP 2020. Raman spectroscopy (LabRAM HR 800, Horiba coupled with an Olympus BX41 optical microscope) was used to collect Raman spectra. The concentration of nitrate ions was analyzed by combustion ion chromatography (ICS 3000, Dionex).

4. Photocatalytic Measurement

The photocatalytic activity of the as-synthesized samples for NO_x removal was evaluated by a Horiba APNA-370 system according to the literature with a bit of modification [10]. The inlet gas flow was a mixture of NO_x gases with an initial concentration of 300 ppb of NO and 30 ppb of NO₂. A photocatalytic paste was prepared by mixing 0.05 g of the photocatalyst in 2 mL ethanol. The

paste was then coated on a dish (diameter, d=5 cm), dried at room temperature and placed in the reactor. A 150 W commercial xenon arc lamp (Abet technologies, LSA-150) was used as a source of light. The experiments were conducted for 30 min in the dark, followed by simulated solar light irradiation for 60 min.

RESULTS AND DISCUSSION

1. X-ray Diffraction Analysis

Fig. 2 shows the XRD patterns of P25, b-TiO₂, n-TiO₂, and n/b-TiO₂. The diffraction peaks of P25 reveal a mixture of anatase (JCPDS No.21-1272) and rutile (JCPDS No.46-1238) phases. In

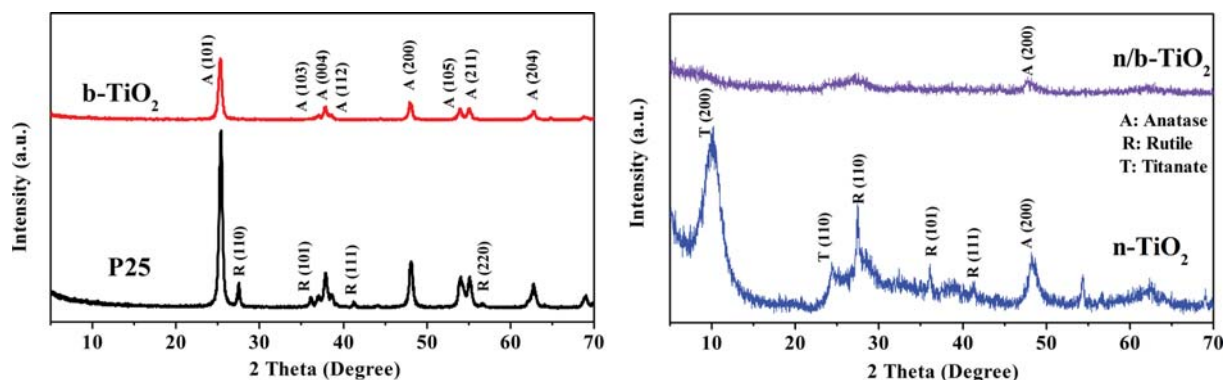


Fig. 2. XRD diffraction patterns of samples.

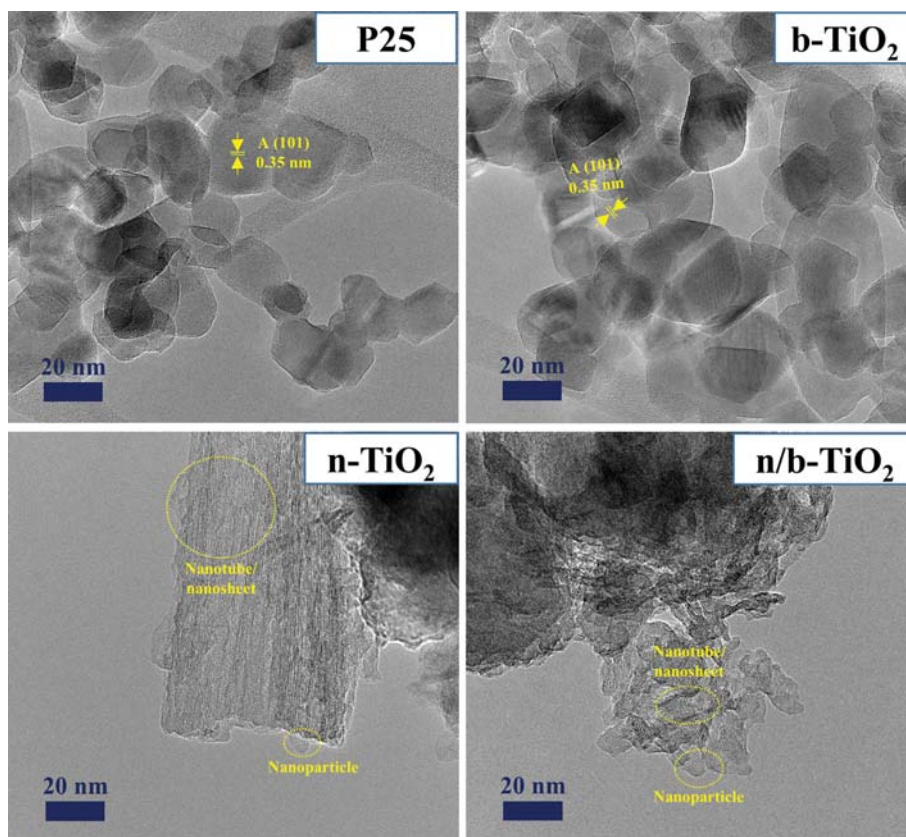


Fig. 3. HRTEM images of samples.

the b-TiO₂, the peaks corresponding to the rutile phase disappear, with just anatase phase remaining. In the nanostructured TiO₂, both anatase and rutile phases are present, which are distinctly represented by the appearance of anatase phase, A (200), and rutile phase, R (110), R (101), and R (111), peaks. However, all diffraction peaks of the rutile phase disappeared in the nanostructured blue TiO₂ (n/b-TiO₂). The absence of the rutile peaks in both b-TiO₂ and n/b-TiO₂ is due to the selective reduction of the rutile phase by Li/EDA reducing agent during the reaction. The possible mechanism to explain the selective reduction of the rutile phase, from a mixture of both rutile and anatase phases, may be due to the different protonation constants of these two phases against pH [11,13,15]. Following that, the anatase phase was stabilized in a super base Li/EDA environment, whereas the rutile phase was damaged easily into the amorphous phase. Due to this selective reduction phenomenon, the surface of b-TiO₂ and n/b-TiO₂ produced a disordered layer. Kim et al. [15] indicated that the disorder shell content has an effect on the coloring of the product, which changes from light blue to black depending on the percentage of the rutile phase in the precursor.

2. TEM Observation

Fig. 3 shows the HRTEM images of P25, b-TiO₂, n-TiO₂, and n/b-TiO₂. The P25 and b-TiO₂ consist of the spherical or oval shape of nanoparticles with a diameter ranging from 10 to 30 nm. The HRTEM image of P25 revealed the existence of the lattice fringes (101) corresponding to the anatase phase. The n-TiO₂ and n/b-TiO₂ are composed of a mixed form morphologies consisting of spherical nanoparticles, nanosheets as well as nanotubes. The presence of different morphological and structural variations could lead to

create heterojunctions. It has been revealed from the studies [18, 19] that the presence of heterojunctions in the photocatalyst enhance the photocatalytic activities by effective separation of photoinduced electrons and holes.

3. Raman Analysis

Fig. 4 shows the Raman spectra of samples. P25 exhibits five peaks of anatase phase located at 148.8, 200.3, 399.0, 516.1, and 638.5 cm⁻¹, which correspond to Eg, Eg, B1g, A1g, and E2g modes [20]. These peaks shift to 147.0, 196.7, 400.8, 519.6, and 642.0 cm⁻¹ in b-TiO₂, respectively. The n-TiO₂ has three main Raman peaks at 281.9, 448.7, and 709.5 cm⁻¹. In the n/b-TiO₂, the peaks centered at 448.7 and 709.5 cm⁻¹ remain at the same position as compared to the corresponding peak position in the n-TiO₂ sample. However, the peak at 281.9 cm⁻¹ is shifted slightly towards the lower wave-number at 280.1 cm⁻¹ in the n/b-TiO₂. The shifting of Raman peaks in the reduced TiO₂ products might be due to the formation of disorder shell after Li/EDA reduction process. The disorder shell might have contributed to decrease in Raman peak intensities in b-TiO₂ and n/b-TiO₂ samples as compared to P25 and n-TiO₂ samples.

4. XPS Analysis

The surface chemistry, binding energy, and the changes of chemical and electronic states of elements of the prepared samples were investigated by XPS analysis. Fig. 5(a)-(d) show the deconvoluted XPS spectra of O 1s state of as-synthesized samples. The results indicate that the O 1s state of non-reduced samples (P25 and n-TiO₂) possess two different broad peaks with variable intensities. The peaks at 530.6 eV (P25) and 530.3 eV (n-TiO₂) can be ascribed to the lattice oxygen (O_L). Similarly, the deconvoluted peaks at 531.7 eV (P25) and 532.7 eV (n-TiO₂) are contributed by non-lattice oxygen or chemisorbed oxygen (O_C). On the other hand, an additional deconvoluted peak that appears at 530.6 eV is assigned to the oxygen vacancy (O_V) in the reduced blue TiO₂ samples, as shown in Fig. 5(c)-(d). The deconvoluted XPS spectra of Ti 2p state of the samples are shown in Fig. 5(e)-(h). Each deconvoluted Ti 2p spectrum of P25 and n-TiO₂ is composed of only two peaks corresponding to Ti⁴⁺ 2p_{3/2} and Ti⁴⁺ 2p_{1/2}. Interestingly, the deconvoluted Ti 2p spectra of b-TiO₂ and n/b-TiO₂ samples exhibit two new peaks at 460.2 and 463.4 eV (b-TiO₂), and 459.6 and 462.4 eV (n/b-TiO₂), representing Ti³⁺ 2p_{3/2}, and Ti³⁺ 2p_{1/2} states, respectively. Here, the experimental results confirm the formation of Ti³⁺ species and oxygen vacancies due to the selective reduction of the rutile phase of TiO₂ by Li/EDA.

5. Photoluminescence Spectra

The PL spectra were measured to analyze the recombination efficiency of photogenerated carriers, which is quantified by PL intensity. As shown in Fig. 6, the lower emission peak intensities in b-TiO₂ and n/b-TiO₂ indicate the lower recombination rate of photoelectrons and holes as compared to P25 and n-TiO₂ samples. The presence of Ti³⁺ species and oxygen vacancies are possible contributing factors to decrease PL intensities in the reduced b-TiO₂ and n/b-TiO₂ samples. These defects act as trapping centers for electrons and holes, thereby enhancing the separation of photogenerated charge carriers. Moreover, PL emission peaks intensities of n/b-TiO₂ are lower than that of b-TiO₂. This outcome is probably due to the formation of heterojunctions among spherical or oval nanoparticles,

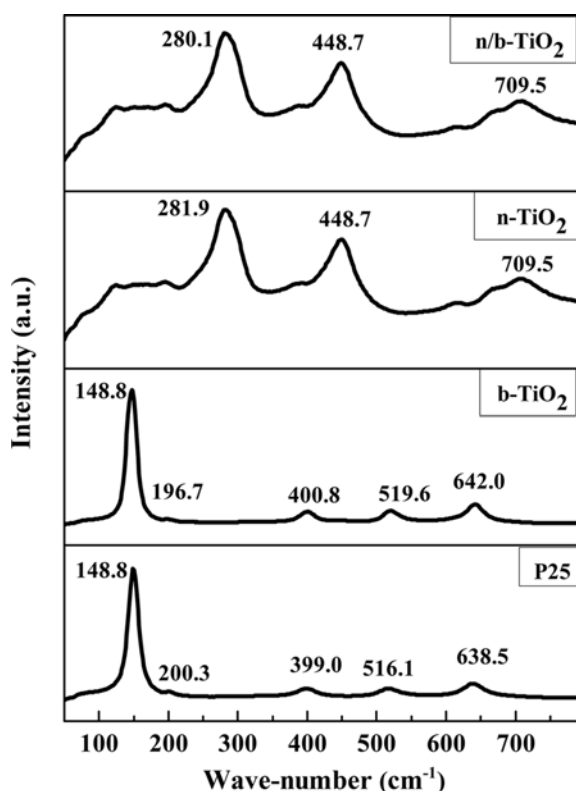


Fig. 4. Raman spectra of samples.

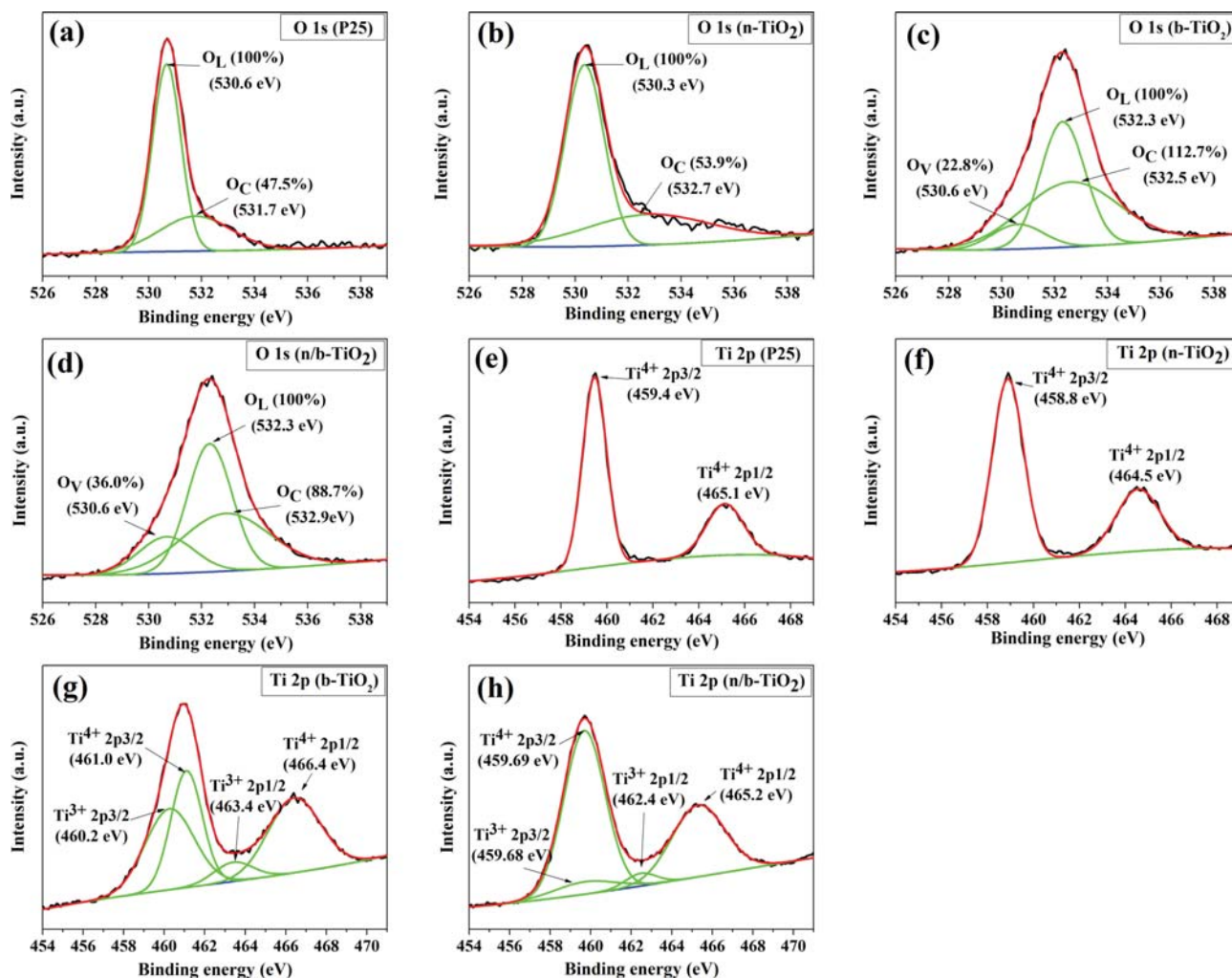


Fig. 5. X-ray photoelectron spectra of samples.

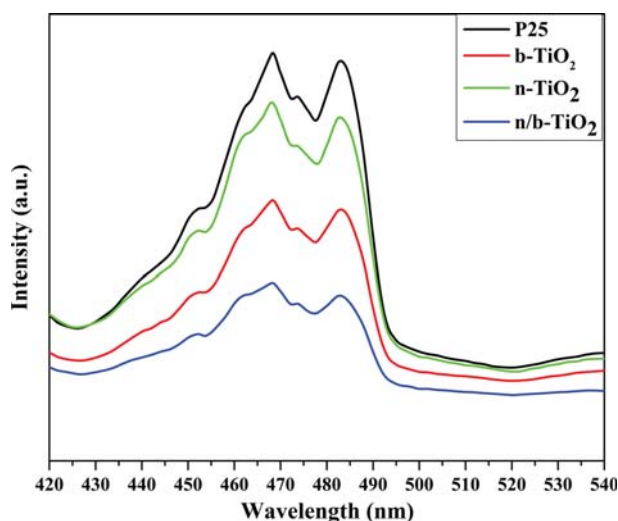


Fig. 6. Photoluminescence spectra of samples.

nanosheets, and nanotubes in n/b-TiO₂. As a result, photogenerated charge carriers spread through heterojunctions leading to efficient

charge separation.

6. UV-Vis DRS Analysis

UV-Vis DRS was measured to investigate the photo-responsive abilities and band-gap energies of the samples, as shown in Fig. 7. The UV-Vis spectra of the reduced samples (b-TiO₂, and n/b-TiO₂) demonstrated a strong absorption in the visible region above 400 nm. However, P25 and n-TiO₂ samples show the light absorption in the UV region only. The existence of superficial defects, including Ti³⁺ and oxygen vacancies in b-TiO₂ and n/b-TiO₂ samples, are ascribed to this outcome. These defects may form the mid-gap states in the band-gap energy, which facilitates photoelectrons to excite and jump into the mid-gap stages even in the low energy visible light range. Moreover, it is interesting that the light absorption in the visible spectrum by b-TiO₂ and n/b-TiO₂ is gradually increased at higher wavelength, as depicted in Fig. 7(a). Particularly, the light absorption spectrum of n/b-TiO₂ indicates that from 600 to 800 nm wavelength, the absorbance is higher than the ultraviolet region. The band gap values of P25, n-TiO₂, b-TiO₂, and n/b-TiO₂, calculated by using the Kubelka-Munk method are 3.28, 3.45, 3.20, and 3.12 eV, respectively. Although the band gap values are relatively higher for the visible light to be absorbed by b-TiO₂ and n/b-TiO₂ samples,

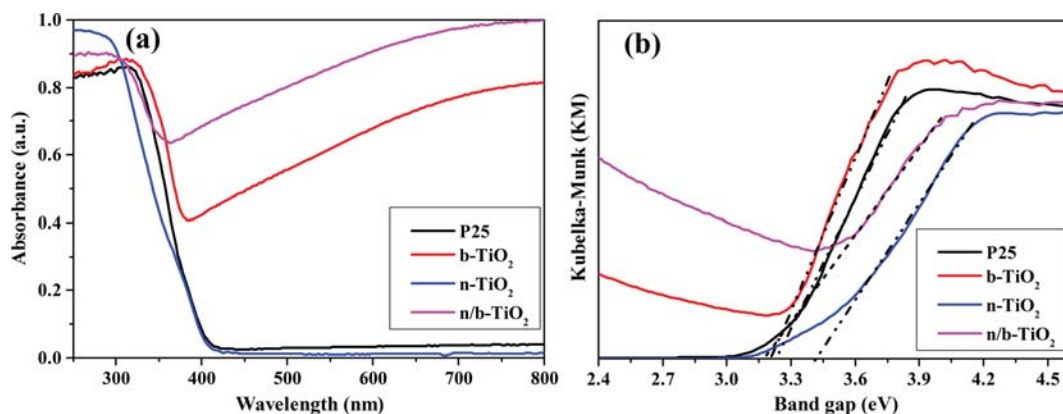


Fig. 7. UV-vis spectra (a) and band gap energy (b) of samples.

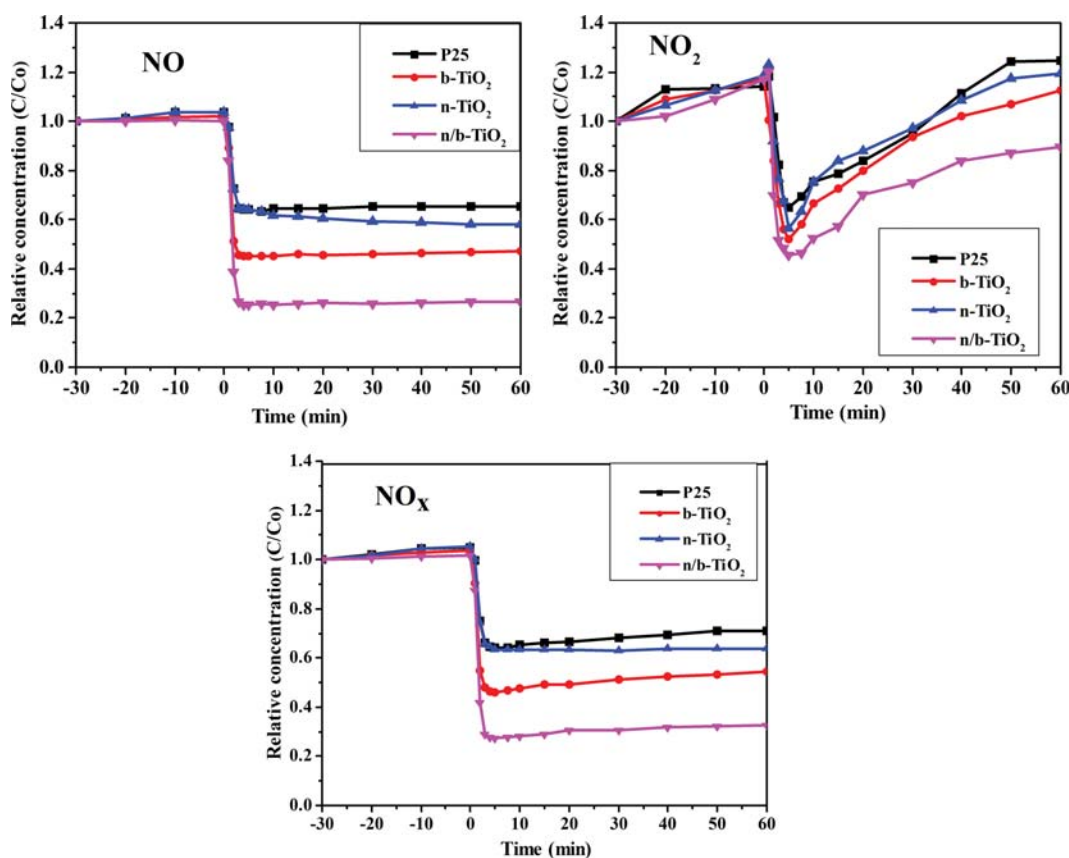


Fig. 8. Photocatalytic removal of NO_x.

the formation of mid-gap stages due to the existence of Ti³⁺ species is responsible for the outcome.

7. Photocatalytic Performance

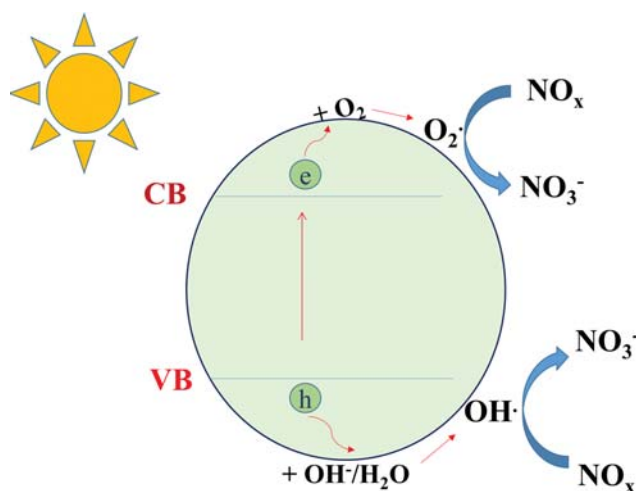
The photocatalytic performance of P25, b-TiO₂, n-TiO₂, and n/b-TiO₂ was evaluated through the transformation of relative concentration (C/C₀) of NO and NO₂. The decrease in concentration of NO_x gases by different samples under simulated solar light is shown in Fig. 8. All samples show a decrease of NO_x concentration after the first 5 min of irradiation. Additionally, the reduced TiO₂ samples, b-TiO₂ and n/b-TiO₂, show better performance for

NO_x removal efficiency, corresponding to 44.8% and 72.6%. Meanwhile, P25 and n-TiO₂ were able to decrease only 36.1% and 37.1% of the initial concentration. The comparative analysis of the NO_x removal efficiency, listed in Table 1, indicates that the material n/b-TiO₂ displays higher potential for the treatment of environmental pollution [21–26].

The possible mechanism is illustrated in Fig. 9. As TiO₂ is exposed with solar light, electrons in the valence band (VB) absorb photons to jump into the conduction band (CB) and leave behind holes. In VB, the holes react with hydroxyl ions (OH⁻) from the water or

Table 1. Comparative study of the NO_x photo-oxidation efficiency of TiO₂ based photocatalysts

Samples	Experiment conditions	NO removal (%)	References
TiO ₂ /C	6 W UV lamp, 365 nm, 200 ppb NO, GHSV=16-98 h ⁻¹	66	[21]
B-PCN	300 W halogen lamp (420-700 nm), 500 ppb NO, 0.2 g Catalyst	44.1	[22]
Ag/TiO ₂	UV and visible light source, gas stream of NO/Ar	70	[23]
g-C ₃ N ₄	Visible light, 500 ppb NO, 0.2 g catalyst	38.6	[24]
N-TiO ₂ /g-C ₃ N ₄	Visible light, 0.2 g catalyst, 600 ppb NO	46.1	[25]
BiNPs-decorated-C ₃ N ₄	Visible light, 500 ppb NO, 0.2 g catalyst	60.8	[24]
Pt-doped TiO ₂	50 mg of catalyzer, uv and visible light, RH=70%, gas of NO, NO ₂ and wet air	30	[26]
n/b-TiO ₂	0.05 g of catalyst, 150 W commercial xenon arc lamp, 300 ppb NO	72.6	Our work

**Fig. 9. Schematic mechanism of NO_x photo-oxidation.**

inorganic substances to form hydroxyl radical (OH[·]), which further combines with NO or NO₂ to produce a final product NO₃⁻ [9,27,28]. In CB, excited electrons reduce adsorbed oxygen to form superoxide (O₂⁻), then the superoxide reacts with NO_x to form NO₃⁻ [1,9,29].

Moreover, the oxidation of NO_x by the photoreactions was supported by the formation of nitrate ions that was analyzed by ion chromatography. The amount of nitrate ions adsorbed on the surface of P25 and TiO₂-p during 1 h of the irradiation corresponds to 0.69 and 3.23 mg/L; however, this number increases to 6.26 and 15.62 mg/L for b-P25 and n/b-TiO₂, respectively.

Clearly, the surface adsorbed substances (oxygen, water, organic chemicals) on the TiO₂ play an important role in photocatalysis because they serve as an electrolyte needed for the redox reaction promoted by the photoelectron-hole pairs [1]. The percentage of O_C on the surface of different TiO₂ samples obtained via deconvolution of XPS is 47.5% (P25), 53.9% (n-TiO₂), 112.7% (b-TiO₂), and 88.7% (n/b-TiO₂). The O_C percentage is consistent with the NO_x abatement induced by P25 (36.1%), n-TiO₂ (37.1%), b-TiO₂ (44.8%), and n/b-TiO₂ (72.6%). The improved photocatalytic activity of n/b-TiO₂ in comparison with b-TiO₂ may be due to the lower recombination rate of photogenerated charge carriers, which is evidenced from the PL spectra. Note that the NO_x removal performance of n/b-TiO₂ is higher than that of b-TiO₂. The possible reason

might be due to the junction effect created from different crystal structures (nanoparticles, nanorods, nanotubes, and nanosheets) which exists in the n/b-TiO₂. According to Jiang et al. [25], the heterojunction might reduce the electron-hole recombination, thereby improving the utilization of these charge particles. In the n/b-TiO₂, the junction phase structure can make the photoelectrons transport among the different CBs of nanosheets, nanotubes, nanorods, and nanoparticles instead of jumping down to VBs and recombining with the holes. Therefore, the lifetime of electrons and holes in n/b-TiO₂ is expected to be prolonged more than the b-TiO₂, leading to enhancing photocatalytic activity. Furthermore, the formation of Ti³⁺ species, oxygen vacancies, at the blue photocatalyst also contributed to the photocatalytic activity of samples, causing better activity of blue TiO₂ than white TiO₂.

CONCLUSION

Blue TiO₂ nanomaterials were synthesized successfully by Li/EDA reduction method. The experimental results indicated the enhancement of light absorption in the visible region of the solar spectrum, reduction of band gap energy, formation of superficial defects such as oxygen vacancies and Ti³⁺ species, and the decrease of the recombination rate of e-h pairs in the reduced blue TiO₂ as compared to white TiO₂ (P25 and n-TiO₂). The experimental results also showed the enhancement of photo-oxidation of NO_x gases using the blue TiO₂ with the highest efficiency of the n/b-TiO₂ (72.6%).

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REFERENCES

1. J. Ângelo, L. Andrade, L. M. Madeira and A. Mendes, *J. Environ. Manage.*, **129**, 522 (2013).
2. D. L. Mauzerall, B. Sultan, N. Kim and D. F. Bradford, *Atmos. Environ.*, **39**, 2851 (2005).
3. A. Chaloulakou, I. Mavroidis and I. Gavril, *Atmos. Environ.*, **42**, 454 (2008).

4. T. Whyand, J. R. Hurst, M. Beckles and M. E. Caplin, *Respir. Res.*, **19**, 79 (2018).
5. B. Yacoubi, L. Samet, J. Bennaceur, A. Lamouchi and R. Chtourou, *Mater. Sci. Semicond. Process.*, **30**, 361 (2015).
6. S. Muneer and J. H. Lee, *Sci. Rep.*, **8**, 12291 (2018).
7. C. Locci, L. Vervisch, B. Farcy, P. Domingo and N. Perret, *Flow Turbulence Combust.*, **100**, 301 (2018).
8. S. Barman and L. Philip, *Environ. Sci. Technol.*, **40**, 1035 (2006).
9. A. Martinez-Oviedo, S. K. Ray, H. P. Nguyen and S. W. Lee, *J. Photochem. Photobiol. A Chem.*, **370**, 18 (2019).
10. H. P. Nguyen, M. Matsuoka, T. H. Kim and S. W. Lee, *J. Photochem. Photobiol. A Chem.*, **367**, 429 (2018).
11. K. Zhang, L. Wang, J. K. Kim, M. Ma, G. Veerappan, C. L. Lee, K. Kong, H. Lee and J. H. Park, *Energy Environ. Sci.*, **9**, 499 (2016).
12. H. H. Nguyen, G. Gyawali, T. H. Kim, S. Bin Humam and S. W. Lee, *Prog. Nat. Sci. Mater. Int.*, **28**, 548 (2018).
13. S. Bin Humam, H. H. Nguyen, C. Regmi, G. Gyawali, B. Joshi and S. W. Lee, *Ceram. Int.*, **45**, 4230 (2019).
14. Y. Yan, B. Hao, D. Wang, G. Chen, E. Markweg, A. Albrecht and P. Schaaf, *J. Mater. Chem. A*, **1**, 14507 (2013).
15. Y. Kim, H. M. Hwang, L. Wang, I. Kim, Y. Yoon and H. Lee, *Sci. Rep.*, **6**, 25212 (2016).
16. J. Yang, C. Zheng, P. Xiong, Y. Li and M. Wei, *J. Mater. Chem. A*, **2**, 19005 (2014).
17. X. Lu, G. Wang, T. Zhai, M. Yu, J. Gan, Y. Tong and Y. Li, *Nano Lett.*, **12**, 1690 (2012).
18. C. N. Ri, S. G. Kim, J. Y. Jong, S. N. Pak, S. C. Ri and J. H. Ri, *New J. Chem.*, **42**, 647 (2018).
19. J. Zhu, J. He, L. Hu and L. Da, *J. Solid State Chem.*, **276**, 104 (2019).
20. H. H. Nguyen, G. Gyawali and S. W. Lee, *J. Ceram. Process. Res.*, **17**, 409 (2016).
21. C. H. Ao and S. C. Lee, *Appl. Catal. B: Environ.*, **44**, 191 (2003).
22. J. Cao, J. Zhang, X. Dong, H. Fu, X. Zhang, X. Lv, Y. Li and G. Jiang, *Appl. Catal. B: Environ.*, **249**, 266 (2019).
23. M. Xu, Y. Wang, J. Geng and D. Jing, *Chem. Eng. J.*, **307**, 181 (2017).
24. G. Jiang, X. Li, M. Lan, T. Shen, X. Lv, F. Dong and S. Zhang, *Appl. Catal. B: Environ.*, **205**, 532 (2017).
25. G. Jiang, J. Cao, M. Chen, X. Zhang and F. Dong, *Appl. Surf. Sci.*, **458**, 77 (2018).
26. Y. Hu, X. Song, S. Jiang and C. Wei, *Chem. Eng. J.*, **274**, 102 (2015).
27. J. S. Dalton, P. A. Janes, N. G. Jones, J. A. Nicholson, K. R. Hallam and G. C. Allen, *Environ. Pollut.*, **120**, 415 (2002).
28. Q. L. Yu and H. J. H. Brouwers, *Appl. Catal. B: Environ.*, **92**, 454 (2009).
29. Y. Hu, X. Song, S. Jiang and C. Wei, *Chem. Eng. J.*, **274**, 102 (2015).